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# Risk Characterization in Health Risk Assessments under CMP

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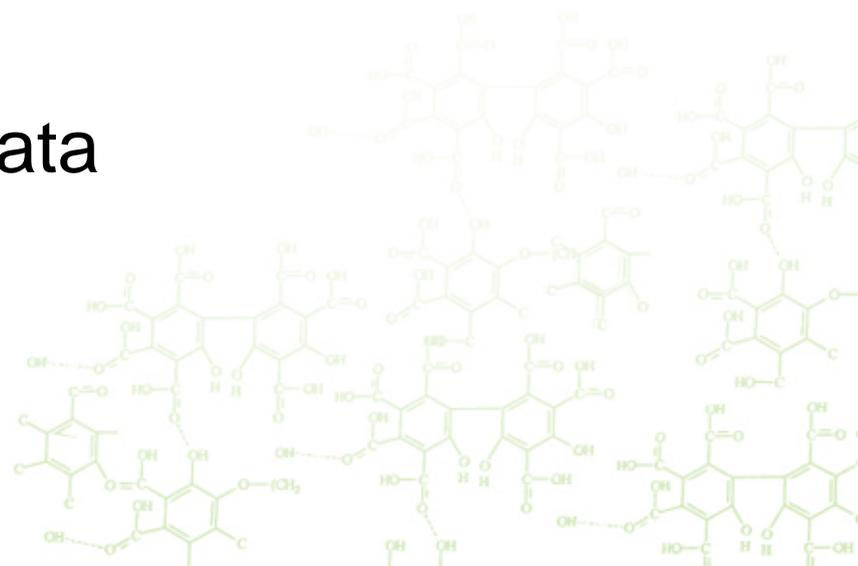
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# Outline

- Alignment with Risk Assessment Toolbox
  - Type 1 Approach
  - Type 2 Approach
  - Type 3 Approach
- Margins of Exposure
- Human Biomonitoring Data



# Risk Assessment Toolbox

## Type 1 Approach

- Addresses the substance/group with a science-based policy response
- Used when regulatory assessment conclusion under s.64 of CEPA 1999 is not suitable
- Examples include: Referring to a better placed program (e.g., foods); documentation of previous action under CEPA 1999

## Type 2 Approach

- Addresses substances using a broad-based approach, often based **on low potential for exposure and conservative scenarios**
- Substances do not meet criteria under s.64
- Examples include: Rapid Screening; Threshold of Toxicological Concern type approaches

Low

Level of Complexity

High

## Type 3 Approach

### Type 3-1

- Addresses the substance/group with a reduced amount of effort for streamlined hazard and/or exposure analysis
- Examples include: Use of international hazard characterizations; use of biomonitoring data; qualitative assessment

### Type 3-2

- Substance/group requires de novo risk assessment

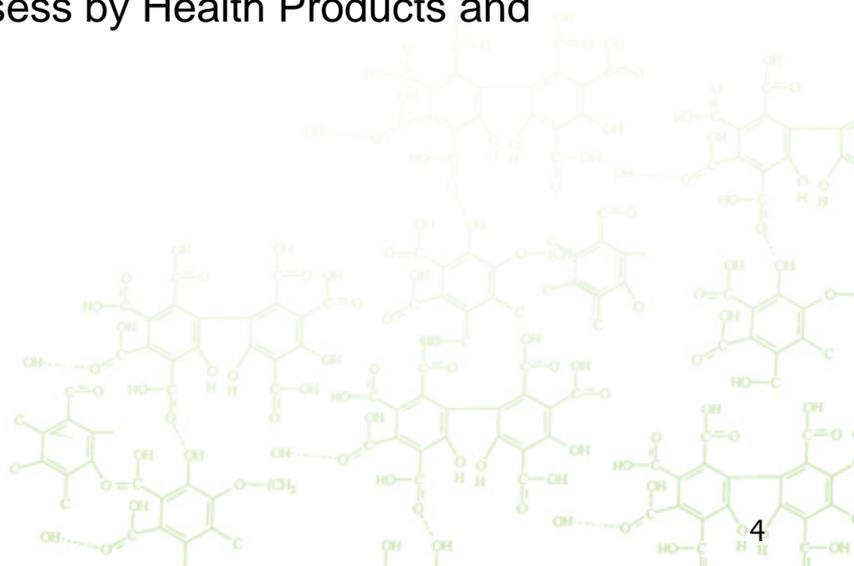
### Type 3-3

- A complex assessment is required for the substance/group that may require cumulative assessment approaches

RM actions for those meeting s.64; additional information gathering and source attribution may be required to inform risk management

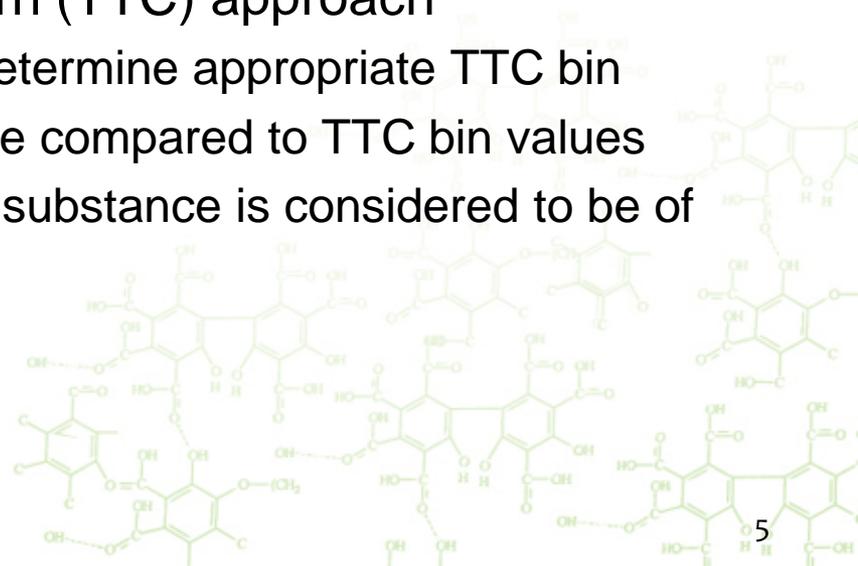
# Type 1 Approach

- Qualitative approach to risk characterization
  - Use of a science-based policy decision or when formal conclusion under S.64 not appropriate
    - e.g., substances addressed under Montreal Protocol
    - Previously addressed under CEPA
    - More appropriately addressed under a different Health Canada program
      - Food related substances assess by Health Products and Foods Branch



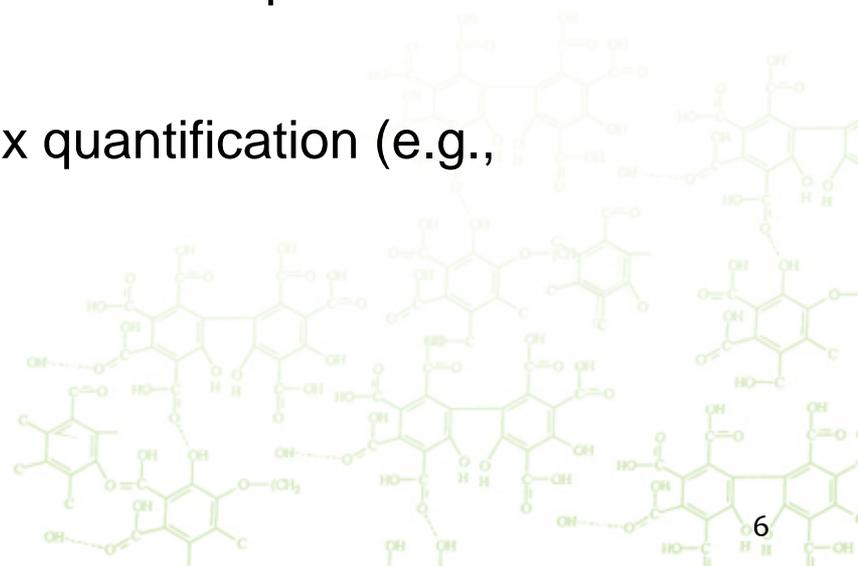
# Type 2 Approach

- Qualitative or semi-quantitative approach to risk characterization
  - Generally adopted for substances for which exposure is anticipated to be very low
  - Substances determined not to be “toxic” under S.64
  - Rapid screening approaches (no direct or indirect exposure sources)
  - Threshold of Toxicological Concern (TTC) approach
    - Screen of health effects data to determine appropriate TTC bin
    - Conservative estimate of exposure compared to TTC bin values
    - If exposure below TTC bin value, substance is considered to be of low concern to health



# Type 3 Approaches

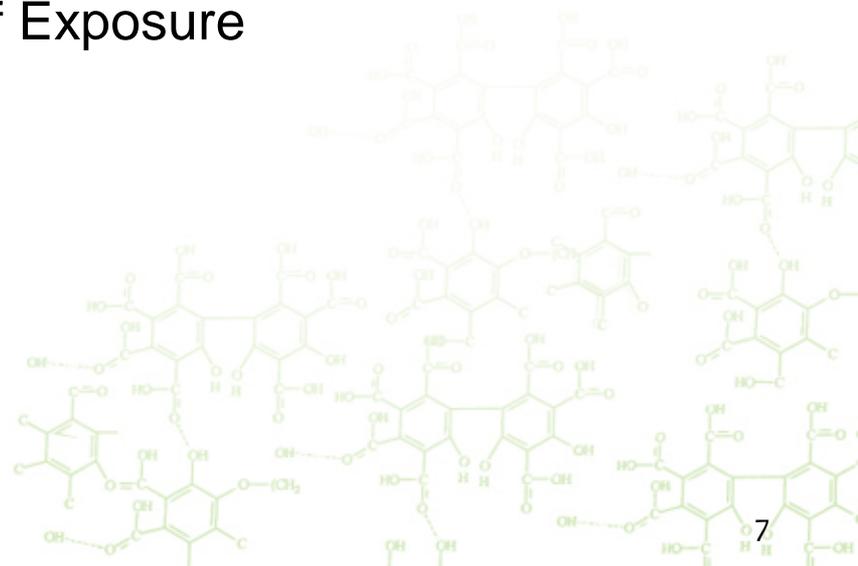
- Quantitative
  - Level of refinement/effort minimum necessary to make a decision
  - Rely on existing information to extent reasonable
    - e.g., use of international hazard classifications as critical endpoint and critical effect levels, with update of literature
    - If no acceptable assessment identified, more in-depth de novo assessment required
  - Quantitative comparison of effect levels to exposure estimates
    - Margin of Exposure approach
  - Sometimes requires more complex quantification (e.g., cumulative assessment)
  - Use of human biomonitoring data



# Margins of Exposure

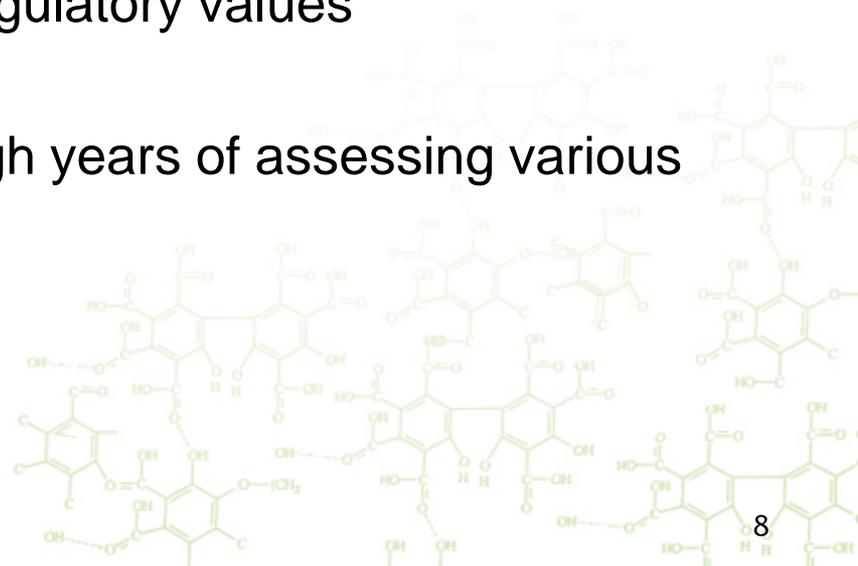
- Comparison of levels of human exposure for different age groups and subpopulations to levels associated with health effects (critical effect levels or Points of Departure: NO(A)EL, LO(A)EL, BMD)

$$\text{MOE} = \frac{\text{Critical Effect Level}}{\text{Estimate of Exposure}}$$



# Interpretation of MOE

- Decision under S.64c based on adequacy of MOE to protect human in light of uncertainties
- **If MOEs don't appear to be adequate, consider further refinement! (iterative process)**
- Decision on adequacy of MOE involves consideration of several factors, including those commonly incorporated in uncertainty or safety factors used in derivation of regulatory values
- Draws on expertise developed through years of assessing various types of substances/datasets



# Interpretation of MoE

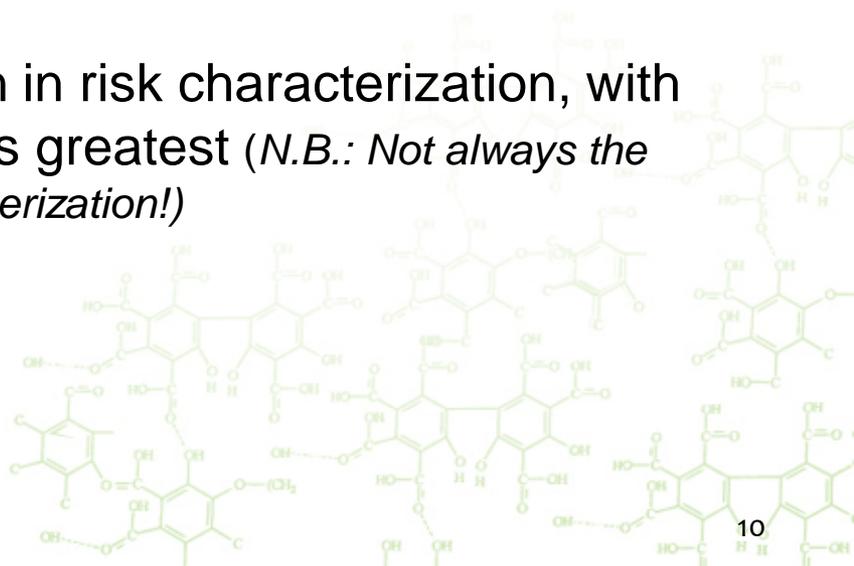
## Factors influencing interpretation of adequacy of MoE:

- Magnitude of margin
- Confidence in databases on effects and exposure
- Interspecies & inter-individual variability in sensitivity (sensitive sub-populations)
- Severity of effect
- Potential relation of critical effect to more severe effects
- Steepness of exposure-response curve
- Dose spacing in critical study
- Existence of lower bound on effect levels
- Potential for exposure from additional sources (concurrent exposures from multiple products)
- Others



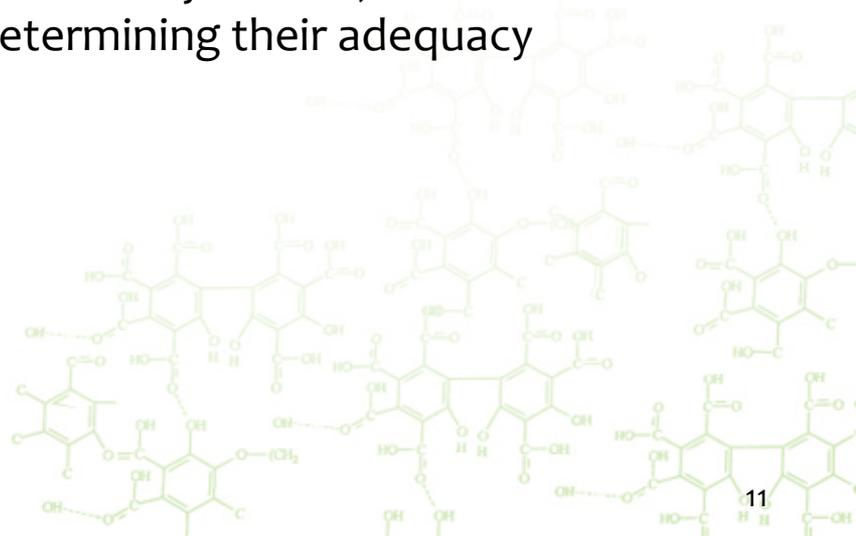
# Multiple MOEs

- MOE are derived for each likely exposure scenario
  - For intermittently used products, short term effect levels compared to shorter term exposure estimates during use of product or daily average estimates
    - E.g., paints, hobbies
  - For longer term frequently used products, longer term/chronic effect levels compared to long term exposures
    - E.g., skin lotion
  - For environmental media, average daily multimedia intake estimates or air concentrations are compared to chronic effect levels
- All MOEs are taken into consideration in risk characterization, with focus on values in which confidence is greatest (*N.B.: Not always the lowest effect level from dose-response characterization!*)
- **Refine as required!**



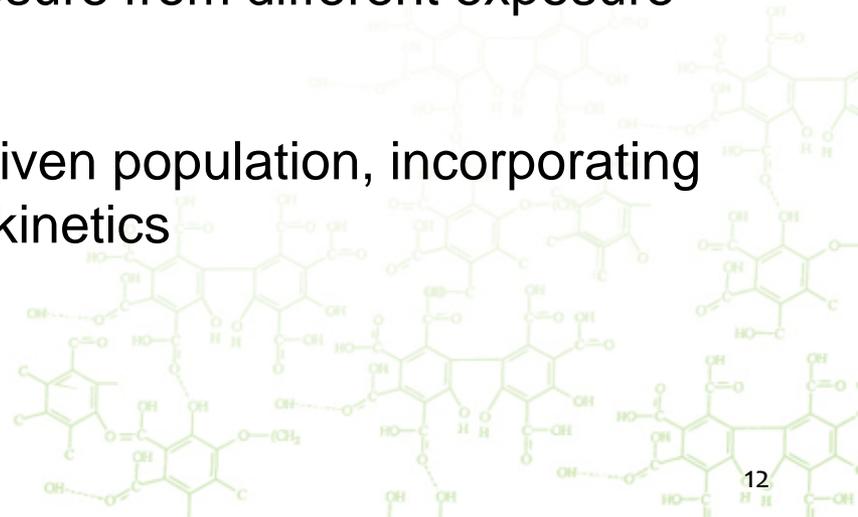
# What MOEs Are Not

- MOEs derived in screening assessments are not:
  - A delineation of “safe” versus “unsafe”
  - An estimate of probability
  - A regulatory guidance value (but related)
    - MOE approach does not use default uncertainty factors or require the development of chemical specific uncertainty factors, but similar information taken into account in determining their adequacy



# Human Biomonitoring Data (HBM)

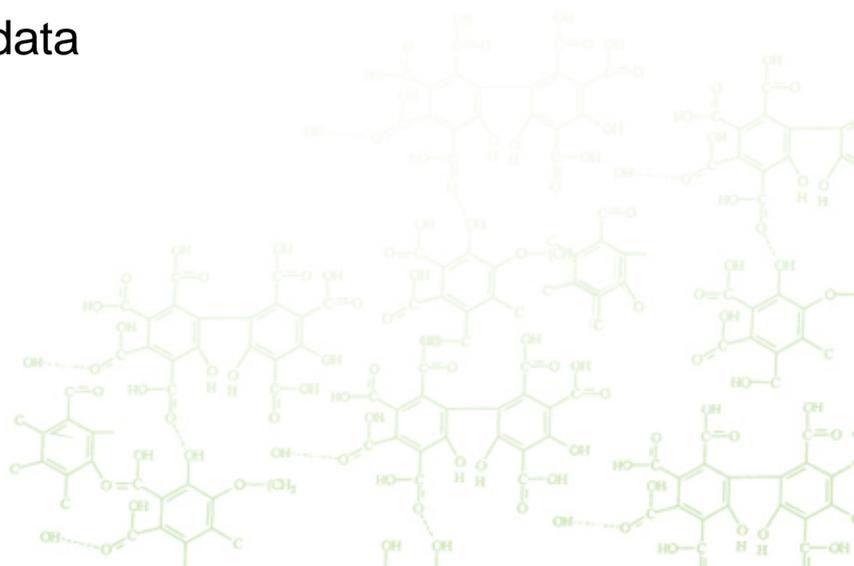
- Amount of human biomonitoring data available increasing rapidly in Canada and elsewhere (e.g., Canadian Health Measures Survey, US National Health and Nutrition Examination Survey (NHANES))
- Chemical substances most commonly measured in breast milk, urine, whole blood and serum
- The use of HBM data in risk assessment allows for direct and a more precise assessment of risk
- Reflective of the absorbed dose into the human body and can provide a measure of integrated exposure from different exposure sources and routes
- Including the distribution of risk in a given population, incorporating individual variability in exposure and kinetics



# Considerations for Use of Human Biomonitoring Data in Risk Assessment

Within the context of CMP, there are a number of considerations prior to incorporation of Human Biomonitoring (HBM) data in human health risk assessment:

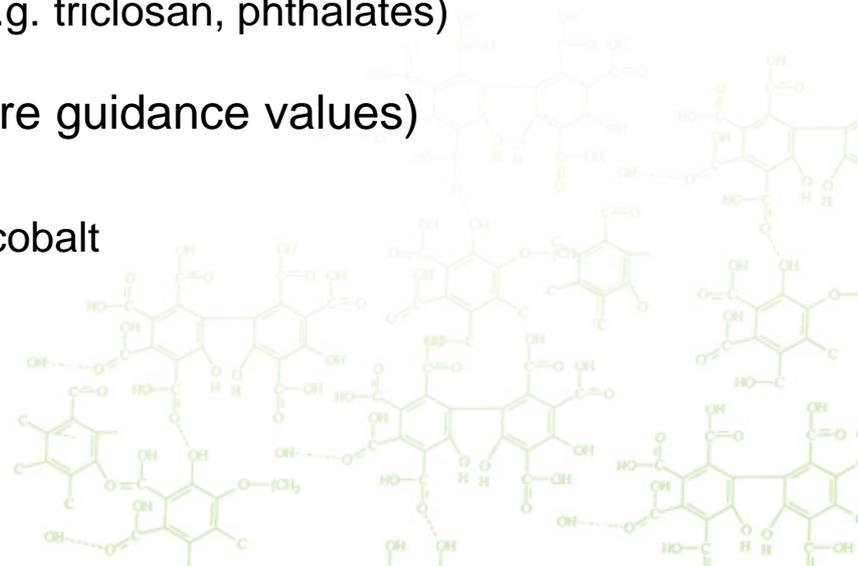
1. Adequacy of the biomarker
2. Quality of the data
3. Appropriateness of the Data Set
4. Approach for interpreting the data



# Use of HBM Data in CMP Risk Assessments

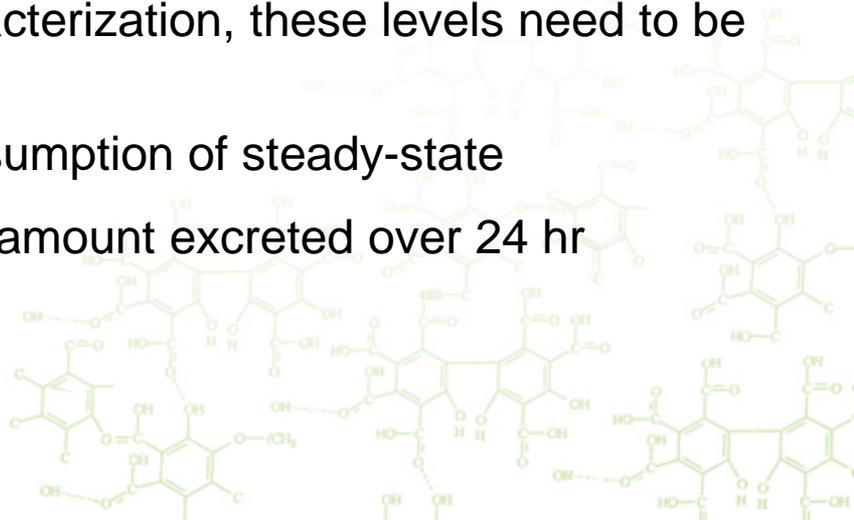
Use of HBM Data has evolved from qualitative to quantitative use including:

- Examining exposure trends and patterns:
  - By sex (e.g. triclosan), age (e.g. PFOA), geography or subpopulations (e.g. selenium), or overall exposure patterns (e.g. cobalt)
- Examining potential association/correlation with health outcomes from cross-sectional health surveys, prospective or retrospective epidemiology studies
  - E.g. Lead (neurodevelopmental); selenium (T2 diabetes)
- Estimating external intakes of exposure
  - Dose-reconstruction or reverse dosimetry (e.g. triclosan, phthalates)
- Comparing with health effects data (exposure guidance values)
  - Directly → lead
  - Indirectly (Forward dosimetry) → selenium; cobalt



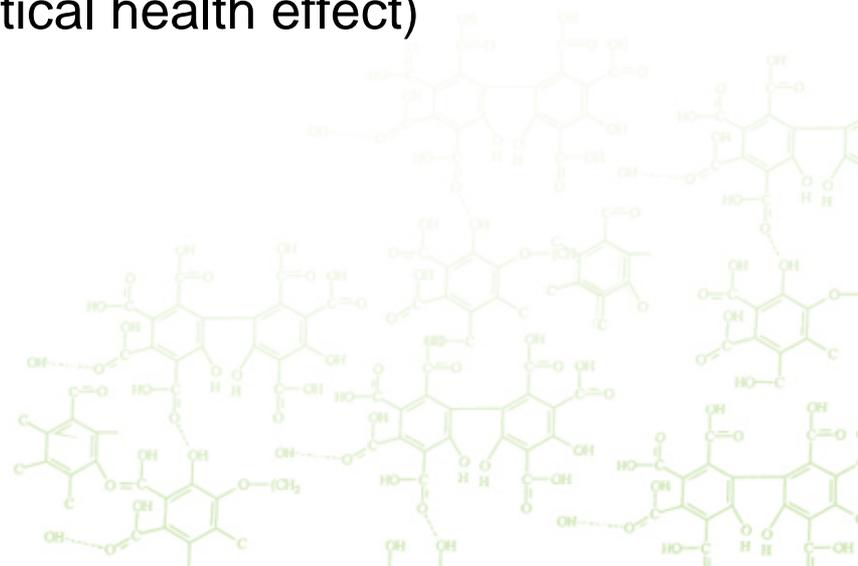
# Uncertainties and Limitations of Using HBM in Risk Assessment

- Not all chemicals are monitored (e.g., issues with sampling techniques)
- The presence of a chemical does not necessarily mean an adverse health effect will occur
- Absence of a chemical does not mean that an exposure did not occur
- HBM data from national surveys alone cannot determine the source or route of exposure
- Relevance & translation of occupational exposure to other populations
- Knowledge of chemical-specific pharmacokinetics and the characteristics of the biomarker as a measure or representative of the external exposure of interest
- Hazard data typically based on intake levels (mg/kg bw/day) vs. internal exposure. For quantitative use in risk characterization, these levels need to be linked.
- There is uncertainty associated with the assumption of steady-state
- Assumptions made to convert spot urine to amount excreted over 24 hr



# Use of HBM Data in Risk Assessment

- Several CMP assessments have used HBM data quantitatively to make conclusions about the potential for risk to human health:
  - PBDEs, HBCD, BPA (use of breastmilk data for estimating dietary intakes of infants)
  - PFOA and PFOS (comparison of blood levels in Canadians with serum levels in rodents from toxicity studies)
  - Lead (whole blood – comparison with neurodevelopmental effects)
  - Cobalt (use of existing biokinetic model studies to derive blood equivalent concentrations to the critical health effect)
  - Triclosan (spot urine)
  - Selenium (whole blood)
  - Phthalates (spot urine)



# Characterization of Uncertainty in CMP Health Assessments

- Describe sources of uncertainty and potential impact on conclusion
  - Interspecies and intraspecies extrapolation (toxicokinetics/dynamics)
  - Uncertainty of analytical measurements
  - Nature or severity of the toxic effect
  - Size/type of population to be protected (sensitive/susceptible populations)
  - Quality of toxicological information
  - Database deficiencies
  - Assumptions related to models
- Identification of Data Gaps and Data Needs
  - Highlights where additional data can help to increase the precision and quality of the decision (reduce uncertainty)
  - Targeted research and monitoring and surveillance initiatives

